Operationalizing the social-ecological systems framework to assess sustainability

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Environmental governance is more effective when the scales of ecological processes are well matched with the human institutions charged with managing human–environment interactions. The social-ecological systems (SESs) framework provides guidance on how to assess the social and ecological dimensions that contribute to sustainable resource use and management, but rarely if ever has been operationalized for multiple localities in a spatially explicit, quantitative manner. Here, we use the case of small-scale fisheries in Baja California Sur, Mexico, to identify distinct SES regions and test key aspects of coupled SESs theory. Regions that exhibit greater potential for social-ecological sustainability in one dimension do not necessarily exhibit it in others, highlighting the importance of integrative, coupled system analyses when implementing spatial planning and other ecosystem-based strategies.


The authors declare no conflict of interest.

This article is a PNAS Direct Submission.

Freely available online through the PNAS open access option.

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This article contains supporting information online at www.pnas.org/lookup/suppl/doi:10.1073/pnas.1414640112/-/DCSupplemental.

Significance

Meeting human needs while sustaining ecosystems and the benefits they provide is a global challenge. Coastal marine systems present a particularly important case, given that >50% of the world’s population lives within 100 km of the coast and fisheries are the primary source of protein for >1 billion people worldwide. Our integrative analysis here yields an understanding of the sustainability of coupled social-ecological systems that is quite distinct from that provided by either the biophysical or the social sciences alone and that illustrates the feasibility and value of operationalizing the social-ecological systems framework for comparative analyses of coupled systems, particularly in data-poor and developing nation settings.

www.pnas.org/cgi/doi/10.1073/pnas.1414640112

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We hypothesized that those regions of BCS with greater potential for social-ecological sustainability in the ecological dimensions (i.e., fish populations and the marine ecosystems they are part of) would exhibit greater potential in the social dimensions (i.e., fishers and the institutions that govern fishers’ interactions with BCS’ marine ecosystems) (SI Appendix, Fig. S1). We also hypothesized that measures of the two social system dimensions (Actors and Governance System) would be positively correlated, as would the measures of the two ecosystem dimensions (Resource System and Resource Units), given the linkages within the social and ecological domains (SI Appendix, Fig. S1). Finally, we predicted that we would observe substantial spatial variation in the potential for social-ecological sustainability.

**Results**

To test these three hypotheses, we first mapped regions of major small-scale fishing activity in BCS. These regions, hereafter referred to as SES regions (Fig. 1), were derived using data from subsequent peer-reviewed literature, government environmental and economic data, and expert knowledge from fishers, resource management and conservation practitioners, and researchers (see SI Appendix for details). This map was essential for the translation of the SES framework and hypothesis tests we report here. If, instead, we had relied on a political map, delineating the five municipalities and/or counties of BCS, the matching of ecological and institutional scales would play a role in determining the likelihood for sustainable governance of fisheries and other common pool resources (e.g., refs. 6, 9, and 10). However, contemporary environmental governance regimes often neglect the array of social, institutional, and ecological factors known to be vital to develop potential for social-ecological sustainability (e.g., refs. 4 and 11–13). We hypothesized that those regions of BCS with greater potential for social-ecological sustainability in the ecological dimensions (i.e., fish populations and the marine ecosystems they are part of) would exhibit greater potential in the social dimensions (i.e., fishers and the institutions that govern fishers’ interactions with BCS’ marine ecosystems) (SI Appendix, Fig. S1). We also hypothesized that measures of the two social system dimensions (Actors and Governance System) would be positively correlated, as would the measures of the two ecosystem dimensions (Resource System and Resource Units), given the linkages within the social and ecological domains (SI Appendix, Fig. S1). Finally, we predicted that we would observe substantial spatial variation in the potential for social-ecological sustainability.

**Table 1. SES variables analyzed for BCS’s small-scale fisheries**

<table>
<thead>
<tr>
<th>Variable</th>
<th>Weight*</th>
</tr>
</thead>
<tbody>
<tr>
<td>Dimension 1: Governance System</td>
<td>1.00</td>
</tr>
<tr>
<td>1. Operational and collective-choice rules</td>
<td>0.50</td>
</tr>
<tr>
<td>2. Territorial use privileges</td>
<td>0.25</td>
</tr>
<tr>
<td>3. Fishing licenses</td>
<td>0.25</td>
</tr>
<tr>
<td>Dimension 2: Actors</td>
<td>1.00</td>
</tr>
<tr>
<td>4. Diversity of relevant actors</td>
<td>0.20</td>
</tr>
<tr>
<td>5. Number of relevant actors</td>
<td>0.20</td>
</tr>
<tr>
<td>6. Migration</td>
<td>0.20</td>
</tr>
<tr>
<td>7. Isolation</td>
<td>0.20</td>
</tr>
<tr>
<td>8. Livelihood diversity potential</td>
<td>0.20</td>
</tr>
<tr>
<td>Dimension 3: Resource Units</td>
<td>1.00</td>
</tr>
<tr>
<td>9. Diversity of targeted taxa</td>
<td>0.50</td>
</tr>
<tr>
<td>10. Per capita revenue</td>
<td>0.50</td>
</tr>
<tr>
<td>Dimension 4: Resource System</td>
<td>1.00</td>
</tr>
<tr>
<td>11. System productivity</td>
<td>0.33</td>
</tr>
<tr>
<td>12. System size</td>
<td>0.33</td>
</tr>
<tr>
<td>13. System predictability</td>
<td>0.33</td>
</tr>
</tbody>
</table>

*Weight refers to the weight given to each lower-tier variable (numbered 1–13), when used to calculate the four first-tier variables (i.e., Dimensions). See Materials and Methods and SI Appendix, section 2 for details.

To operationalize the SES framework for our focal system, we identified 13 variables that have been linked to the likelihood of the emergence of locally appropriate governance of SESs, and small-scale fisheries SESs in particular (19). These 13 variables were nested underneath the four dimensions introduced earlier (Table 1). We then identified indicators for each of the 13 variables and quantified them on the basis of primary data (Table 2). Once we calculated indicators for all 13 variables on a common scale and then created composite, quantitative measures of each of the four SES dimensions (i.e., first-tier variables), we were able to test our hypotheses of social-ecological alignment, within-domain correlation, and spatial variation in the potential for social-ecological sustainability. Fig. 2 provides a visualization of our methods.

Contrary to the first hypothesis, we found few consistent positive relationships between the social and ecological dimensions related to the potential for sustainable resource use (Fig. 3 and SI Appendix, Results). Among the a priori tests we conducted regarding the first-tier SES variables, only one pair exhibited the predicted relationship. Regions characterized by high Governance System scores also had high Resource Units scores (Fig. 3; linear regression: $R^2 = 0.33; F_{1,10} = 4.86; P = 0.05$). This association was particularly evident for regions with the highest and lowest sets of scores: Pacifico Norte and Todos Santos and Cabo San Lucas, Gulf of Ulloa and East Cape, respectively (Fig. 4).
Neither the two first-tier social system variables (Actors and Governance System) nor the two first-tier ecosystem variables (Resource Units and Resource System) were associated, contrary to our second hypothesis (Fig. 3; $P > 0.10$). However, these analyses revealed the dimensions within which the potential for sustainable resource use and governance is particularly high or low, which could inform future capacity-building efforts and other policy and management interventions. For example, although El Corredor’s Governance System score was the lowest of the 12 regions, its Actors score was almost as high as that of Pacífico Norte, indicating that El Corredor already exhibits substantial potential for sustainable resource use in the latter dimension (Fig. 4 and SI Appendix, Table S4).

Finally, the potential for social-ecological sustainability varied substantially among the SES regions, as predicted (Fig. 4). As reported earlier, regions that scored high in one dimension (i.e., one first-tier variable) did not necessarily score high in all four dimensions. Magdalena Bay and Gulf of Ulloa, for example, had Resource Units scores close to 1 (SI Appendix, Table S4), yet the Actors scores for these two regions were both less than the median. Cabo San Lucas, East Cape, La Paz, and Loreto had some of the lowest scores overall (linear contrast of these four regions vs. the other eight, following ANOVA of scores by region: $F_{1, 36 } = 13.08; P = 0.001$). Principal components analysis provided another means of visualizing spatial variation among the regions, suggesting there are multiple paths to achieving sustainability (SI Appendix, Fig. S3). The results of the first-tier variable analyses were also reflected in the primary data (Table 2 and SI Appendix, Table S2). Together, these analyses illustrate substantial spatial heterogeneity in the potential for sustainable resource use related to small-scale fisheries in BCS and also elucidate how this variation is created by a combination of social and ecological factors (SI Appendix, Figs. S4–S8).

### Discussion

Our approach illustrates how diverse qualitative and quantitative datasets can be integrated in a robust and spatially explicit manner to describe multiple SESs and to test related theory. Analyses of the theoretically grounded measures we created (Table 1 and SI Appendix, Tables S2–S4) revealed that regions that are strong in one dimension are not necessarily strong in the other three (Figs. 3 and 4 and SI Appendix, Table S3). Moreover, variation in the potential for social-ecological sustainability exists at a finer spatial scale than that at which the state currently regulates small-scale fisheries (as described in detail in the SI Appendix).

Our translation of the SES framework also highlights how assessments based on solely biophysical or social data may lead to quite divergent conclusions. Consider, for example, Magdalena Bay, where fishers report the most taxon-rich catches of the 12 SES regions. Previous theory and empirical work suggest that such ecological diversity should buffer the coupled SES from disturbances and confer resilience in the face of environmental and institutional changes (20). However, although Magdalena Bay had a very high Resource Units score, its Actors score was among the lowest. So, depending on which type of data one musters regarding the potential for sustainable fisheries, Magdalena Bay could be scored as either well-endowed or quite weak. Perhaps more important, this result suggests that in the Actors dimension, there is opportunity to build management capacity (e.g., by increasing the ratio of permitted to illegal fishers), whereas in the Resource Units dimension, it may be more important to maintain existing management capacity (e.g., by creating institutions to help ensure continued diversity of targeted taxa). These scores are consistent with our personal experiences in this particular region, where the sheer number of fishers, including many unpermitted fishers, and the diversity of gear types and interests they represent contribute to significant social conflict.

### Table 2. Representative data used to calculate the scores for the four first-tier SES variables

<table>
<thead>
<tr>
<th>SES region</th>
<th>Index, local rules</th>
<th>Total number of fishers reported</th>
<th>Per capita revenue, $USD</th>
<th>Mean chlorophyll $a$, $\mu g \cdot m^{-3}$</th>
<th>CV (coefficient of variation), mean $chl a$</th>
</tr>
</thead>
<tbody>
<tr>
<td>1. Guerrero Negro</td>
<td>0.42</td>
<td>293</td>
<td>39</td>
<td>12,434</td>
<td>2.73</td>
</tr>
<tr>
<td>2. Pacífico Norte</td>
<td>1.00</td>
<td>1,082</td>
<td>73</td>
<td>15,123</td>
<td>1.76</td>
</tr>
<tr>
<td>3. Gulf of Ulloa</td>
<td>0.25</td>
<td>614</td>
<td>85</td>
<td>14,467</td>
<td>2.64</td>
</tr>
<tr>
<td>4. Magdalena Bay</td>
<td>0.25</td>
<td>1,283</td>
<td>90</td>
<td>15,060</td>
<td>2.19</td>
</tr>
<tr>
<td>5. Todos Santos</td>
<td>1.00</td>
<td>77</td>
<td>59</td>
<td>22,243</td>
<td>1.13</td>
</tr>
<tr>
<td>6. Cabo San Lucas</td>
<td>0.25</td>
<td>81</td>
<td>9</td>
<td>2,337</td>
<td>0.84</td>
</tr>
<tr>
<td>7. East Cape</td>
<td>0.25</td>
<td>247</td>
<td>36</td>
<td>2,641</td>
<td>0.88</td>
</tr>
<tr>
<td>8. La Paz</td>
<td>0.25</td>
<td>974</td>
<td>55</td>
<td>986</td>
<td>1.22</td>
</tr>
<tr>
<td>9. El Corredor</td>
<td>0.25</td>
<td>102</td>
<td>52</td>
<td>8,320</td>
<td>1.15</td>
</tr>
<tr>
<td>10. Loreto</td>
<td>0.58</td>
<td>152</td>
<td>44</td>
<td>5,220</td>
<td>1.43</td>
</tr>
<tr>
<td>11. Mulegé</td>
<td>0.58</td>
<td>126</td>
<td>54</td>
<td>6,750</td>
<td>2.31</td>
</tr>
<tr>
<td>12. Santa Rosalia</td>
<td>0.25</td>
<td>523</td>
<td>67</td>
<td>9,803</td>
<td>1.50</td>
</tr>
</tbody>
</table>

The full dataset can be found in SI Appendix, Table S2.

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**Step 1. Identify SES variables relevant to the focal system(s)**

(Table 1 and SI Appendix, Table S1).

**Step 2. Create theoretically grounded indicators for the variables, based on primary data (SI Sub-Appendix D).**

**Step 3. Document relevant trends in primary data (Table 2 and SI Appendix, Table S2).**

**Step 4. Calculate indicator rankings from primary data (Table S3).**

**Step 5. Calculate variable scores (Table S4).**

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Fig. 2. Steps to translate the SES framework into quantitative measures of the potential for social-ecological sustainability, with references to the relevant SI Appendix sections.
These results demonstrate how integrative, interdisciplinary research that includes both qualitative and quantitative data may be synthesized to yield a richer understanding of coupled SESs in particular places. Many of these data have never before been mapped at this scale, and yet such fine-scale information could help inform implementation of ecosystem-based fisheries management, marine spatial planning, and related strategies. The variation among the SES regions suggests that certain marine management strategies, implemented at particular geographic scales, are likely to be more effective in some places than others (see also refs. 21 and 22).

Mapping is necessarily a political project in which local actors must be involved to negotiate boundaries in ways that are more likely to lead to just outcomes (23, 24). Our purpose here is not to portray the SES regions as definitive boundaries but, rather, to encourage finer-scale, spatially explicit governance of SESs, in which local stakeholders participate in the necessary refinement of social-ecological management boundaries. Stakeholder engagement will require and also provide an impetus to create finer-resolution spatial data at the level of individual fishing communities. To our knowledge, these are not yet available for most of BCS and other coastal marine SESs around the world.

To extend our approach and to evaluate spatial heterogeneity in the actual interactions and outcomes associated with small-scale fisheries will require substantially more and different types of data than those currently available.

The diversity of species and fishing practices that are inherently part of small-scale fisheries like BCS’ present other challenges for coupled systems analyses, and for the SES framework specifically. By design, the framework focuses on interactions between resource users and other actors regarding a specific resource in the context of a particular SES. However, small-scale fishers target tens, if not hundreds, of species, which vary in their life histories, economic value, and many other important characteristics. We managed this complexity by scaling up our analysis to the level of major fishing areas, which represented fishers’ use of ocean space to catch many species over the entire year (Fig. 1). Nonetheless, this level of analysis obscures some valuable information.

Similarly, although we can think of biogeographic, oceanographic, and human history as exogenous drivers of the dynamics of a given SES region, on long time scales, they are not static. Global climate change continues to alter the biophysical template on which social-ecological interactions play out; for example, by changing water and air temperatures and the frequency and magnitude of precipitation and coastal storms. Sociopolitical dynamics enter the SES framework both as context and in relationships between attributes internal to the coupled system (25). Evolving sociopolitical dynamics from local to national scales and their interactions with narco-trafficking and other multinational influences shape the opportunities and constraints facing BCS’ small-scale fishers and their decisions about how, where, and when to fish (as in ref. 26).

Our analyses inform four types of management strategies, focused on each of the four dimensions (or first-tier SES variables). Interventions focused on improving existing institutional arrangements are most likely to strengthen the Governance System dimension, whereas those focused on improving relationships among stakeholders will aid in building capacity in the Actors dimension. Similarly, we anticipate that strategies focused on maintaining or improving ecosystem health would benefit the Resource System dimension, and that interventions focused on improving the status of populations of the target species would build capacity in the Resource Units dimension. Such tailored strategies will likely reduce the costs associated with blueprint, or one-size-fits-all, types of policies. Given the physical isolation and consequent high reliance on marine resources in the Pacífico Norte and El Corredor regions, for example, especially when contrasted with the highly populated southern regions, a blueprint approach to fisheries management and MPA implementation will not serve either the human communities or the marine ecosystems of BCS well. Instead, we advocate for more strategic approaches, targeted to the needs and strengths of specific regions. However, it also is important to acknowledge that governance approaches tailored to address problems in one dimension or region could trigger unintended consequences in other dimensions or regions if issues are not addressed holistically.

Applying integrative, place-based understanding of SESs in this and other geographies will enable sustainability science to more fully inform sustainability practice.

Materials and Methods

SES Mapping. To map the SES regions, we began by listing all the small-scale fishing communities along the coast of BCS, with reference to ref. 27. We then identified distinct clusters among them based on four primary factors: biophysical context, including coastal topography, habitats, and species distributions; historical and contemporary coastal land and marine resource use; municipal and state political boundaries; and the concentrations and movement of fishers and fisheries products. Further detail on these factors and the current fisheries management regime can be found in the SI Appendix, Sub-Appendix A.

Based on these first two steps, we drew an initial map of major fishing areas of BCS. This initial map informed a series of unstructured interviews with key informants about the scale and nature of small-scale fisheries activities throughout BCS. Fourth, following refinement of the maps based on the results of the interviews, we created a survey instrument to elicit standardized area-specific expert knowledge of the human and environmental dimensions of BCS’ small-scale fisheries from fisheries managers and conservation and community development practitioners. The survey was distributed to 15 individuals, as included as SI Appendix, Sub-Appendix B. After analysis of the survey results, the SES regions were amended as needed. Further details on the mapping protocol can be found in SI Appendix, Materials and Methods, along with detailed descriptions of
each region (SI Appendix, Sub-Appendix C). The SES regions map and all other maps were produced using ESRI ArcGIS 10.1.

Operationalizing the Framework. Once we had an appropriate map with which to test our hypotheses, we translated the SES framework from a conceptual model into a quantifiable set of indicators, linked with each of the four SES dimensions or first-tier variables. The 13 lower-tier variable selection process was driven by our knowledge of the BCS SESs, review of the scholarly literature, and theoretical underpinnings of our work, including the governance of common-pool resources and the relationships between biodiversity and ecosystem functioning. Fig. 2 provides a step-by-step visualization of the full methods.

Variable Selection and Ranking of the Indicators. A varying number of second, third, and fourth-tier variables are nested underneath each of the four dimensions (Governance System, Actors, Resource System, and Resource Units); in total, there are 13 second-, third-, and fourth-tier variables (Table 1 and SI Appendix, Table S1, after refs. 4 and 19). Importantly, all 13 variables, regardless of whether the primary data used to develop the indicator were qualitative or quantitative, were normalized to a scale of 0–1, so they could be combined and compared. For those variables for which the primary data were continuous, such as total fisheries biomass, the values for each SES region were calculated on the basis of the quantile distribution of the original data. For those variables for which the primary data were qualitative (e.g., presence/absence), they were translated into categorical 0/1 variables. These data and the resulting rankings are presented in SI Appendix, Tables S2–S4.

SI Appendix, Sub-Appendix D includes the description of all 13 SES variables and the related indicators. The description includes each variable’s name, position in the original SES framework (4), definition, theoretical importance, a brief description of the indicator or indicators associated with the variable, and the ranking system for each indicator. Where appropriate, we include the quantile distribution of the original data. For each variable, we identified one or more indicators that enabled either qualitative or quantitative comparison among the 12 SES regions. Together, these complementary indicators captured multiple dimensions of a variable. We developed the ranking system from relevant theory regarding how each variable has been associated with a SES’s potential for sustainable resource governance.

Variable Weighting. Each of the four dimensions, Governance System, Actors, Resource Units, and Resource System, has a cumulative weight score of one (Table 1). The relative contribution of each of the lower-tier variables to this weight depends on the total number of such variables analyzed under a particular dimension, or first-tier variable. For example, the dimension Actors is composed of five lower-tier variables, each of which is weighted 0.20 for an overall score of 1.00 (Table 1 and SI Appendix, Table S1 and SI Appendix, Sub-Appendix D). For the variables that have more than one indicator (SI Appendix, Table S2 and SI Appendix, Sub-Appendix D), scores were averaged before being weighted.

Data and Analyses. The data can be found in SI Appendix, Tables S2–S4. All statistical tests were performed using JMP 11 (SAS Institute) or SPSS 22.0 (SPSS Statistics Inc.). We used standard statistical approaches (i.e., linear regression, analysis of variance, and principal component analysis models) to explore the primary data used to develop the indicators for the SES variables and test a priori hypotheses regarding the first-tier SES variables. Before analysis, the primary data and the calculated variables were plotted to investigate their fit to statistical assumptions; for example, normality. When necessary, data were transformed. The details of these models are described in SI Appendix, Results.

ACKNOWLEDGMENTS. We appreciate the thoughtful and constructive comments provided by S. Levin, M. Ballesteros, and two anonymous reviewers on earlier versions of this manuscript, and we thank our community partners, who have contributed to our individual and collaborative activities in the region. Funding was provided by the US National Science Foundation (GEO 1114964, to H.M.L., S.N., S.M.W.R., and O.A.-O.), The David and Lucile Packard Foundation (H.M.L., O.A.-O., B.E.E., and M.M.-B.), Brown University’s Environmental Change Initiative (H.M.L., L.S., S.N., and S.M.W.R.), Voss Environmental Fellowship Program (K.S.), The Walton Family Foundation (K.B., O.A.-O., B.E.E., M.M.-B., and M.N.), and the World Wildlife Fund Fuller Fellowship Program (M.N.).